# Annual dynamics of $N_2O$ emissions from a tea field in southern subtropical China

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### **ABSTRACT**

The annual dynamics of  $N_2O$  emissions from a tea field in southern subtropical China was observed *in situ* weekly in 2010 using a static closed chamber – gas chromatography (GC) method for three treatments: non-fertilised (CK), conventional (CON) and rice straw mulching (SM). The annual  $N_2O$  emissions for CK, CON and SM were 7.1, 17.2 and 16.7 kg N/ha/year, respectively. The  $N_2O$  emission factors for the CON and SM treatments were estimated as 2.23% and 1.91% of the total fertiliser N applied, respectively. Rice straw mulching exhibited a potential to reduce the  $N_2O$  emissions from the tea field, but not statistically significant (P = 0.82). The daily  $N_2O$  fluxes were positively correlated with the air temperature. The cumulative precipitation of the previous five days was significantly correlated with the daily  $N_2O$  fluxes of CON. The soil water contents were significantly correlated with the daily  $N_2O$  fluxes in the three treatments. The  $N_2O$  fluxes from CON had a more significant correlation with the soil  $NH_4^+$ -N contents than with the soil  $NO_3^-$ -N contents, while the  $N_2O$  fluxes from SM showed an inverse pattern. The  $N_2O$  fluxes from CK did not show any significant relationship with the soil mineral N content.

Keywords: fertilised tea field; nitrous oxide; straw mulching; environmental factors

Nitrous oxide (N<sub>2</sub>O) is a potent greenhouse gas, and its concentration has increased by 18.5% compared with the pre-industrial era and reached 319 ppb in 2005 (IPCC 2007). Soils are considered to be the most important source of N2O emissions (Mosier et al. 1998), and the contribution of agricultural soils to the global N<sub>2</sub>O source is approximately 35% (FAO 2001) because of the excessive use of chemical nitrogen (N) fertilisers. Large amounts of chemical fertilisers applying to croplands induce strong soil acidification and result in very low pH of soils (Guo et al. 2010). The tea tree is a very special plant, and its plantation causes a decline in soil pH, i.e., soil acidification (Han et al. 2007). Nitrate can accumulate in soils even in a very low pH environment, and nitrifying bacteria can adapt to highly acidic soils (Boer and Kowalchuk 2001). A low pH was demonstrated to be associated with a large source of N<sub>2</sub>O from soil nitrification process. Tokuda and Hayatsu (2001) evaluated the potentials of 21 tea soils for N<sub>2</sub>O emissions and found that a negative exponential relationship existed between

the pH of these soils and their potentials for N<sub>2</sub>O emissions and that the N fertiliser application rate of more than 1000 kg N/ha/year significantly enhanced N<sub>2</sub>O emissions from these acidic tea fields. The incorporation of plant residues can also enhance soil N<sub>2</sub>O emissions. Liu et al. (2011) found that the application of wheat straw to the soil increased N<sub>2</sub>O emissions, whereas the incorporation of maize straw did not influence N<sub>2</sub>O emissions, compared to a treatment of chemical fertiliser only. The effect of rice straw mulching on N<sub>2</sub>O emissions in tea fields has not yet been fully explored to determine whether it is an effective practice for improving the environment. Given the boom of the tea industry in southern subtropical China, we monitored the  $N_2O$  emissions from a tea field in the hilly red-soil region maintained under three management practices over the course of an entire year and then analysed the relationships between the N<sub>2</sub>O emissions and environmental factors, aiming to estimate the N<sub>2</sub>O emission factor for tea plantation in the region.

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#### MATERIAL AND METHODS

Site description and treatments. The experimental site is located in a tea field at the Changsha Environmental Observation Station of the Chinese Academy of Sciences (CAS) in Hunan, P.R.China (113°20'E, 28°35'N, elevation of 105 m a.s.l.). 70% of the annual precipitation falls during the warm season from April to June. The soil in the tea field is a Haplic Alfisol, developed from highlyweathered granite parent material, and has a sandy loam texture (clay content of ~25%) and a pH of 4.5. The contents of soil organic carbon (SOC), total soil N, total soil phosphorous, and total soil potassium of the topsoil (0-20 cm) are 11.1, 0.86, 0.37 and 19.67 g/kg soil, respectively. A typical row of tea trees (Figure 1) is characterised by a 1.0 m wide and 0.9 m high canopy (two tea plants grown inside) and a 0.5 m wide inter-row space.

The field experiment was block-designed, carried out in 2010 and included three management practices: non-fertilised (CK, no fertiliser or rice straw application); conventional (CON, fertiliser applied at a rate of 450 kg N/ha/year as two additions: one of 300 kg N/ha as urea on April 15, and one of 150 kg N/ha of oilseed residues, banded 10–15 cm under the soil surface at the fertilisation point shown in Figure 1 on November 12); and rice straw mulching (SM, rice straw at a rate of 4200 kg/ha/year was placed on the soil surface, and the fertiliser application was the same as for CON).

Each treatment had an area of 600 m $^2$  (20 m  $\times$  30 m). The first straw mulching took place on March 25 and the second on November 12.

Nitrous oxide measurement. The N<sub>2</sub>O emissions were measured using a static closed chamber - gas chromatography (GC) method. Each tea row had four specific locations: one inter-row space, one fertilisation point, two under-tree spaces and the in-tree row space. The chamber (0.8 m long  $\times$  0.8 m wide × 1.2 m high) was placed on the soil to cover half of the inter-row space, the in-tree row space, the whole fertilisation point, and one under-tree space to obtain measurements representing the tea field landscape (Figure 1), and three replicates were set for each treatment. The observation period started in January 2010, gas samplings were performed almost once a week between 9:00 and 11:00 am, and the daily N<sub>2</sub>O fluxes were calculated according to Liu et al. (2011). Gas samples were analysed on a GC fitted with a <sup>63</sup>Ni-electron capture detector (Agilent 7890A, Agilent, CA, USA). The annual amount of N<sub>2</sub>O emissions (kg N/ha/year) was estimated by the following equation:

$$\left(\frac{f_1 + f_n}{2} + \sum_{i=1}^{n-1} \frac{(f_{i+1} + f_i) \times (DOY_{i+1} - DOY_i)}{2}\right) \times \frac{365}{DOY_n - DOY_1 + 1} \times \frac{10000}{1000000} \tag{1}$$

where: f – the daily  $N_2O$  flux (mg N/m²/day); the subscripts of 1...n represent the number of the discrete daily  $N_2O$  flux; and DOY denotes the day of the year (1–365).

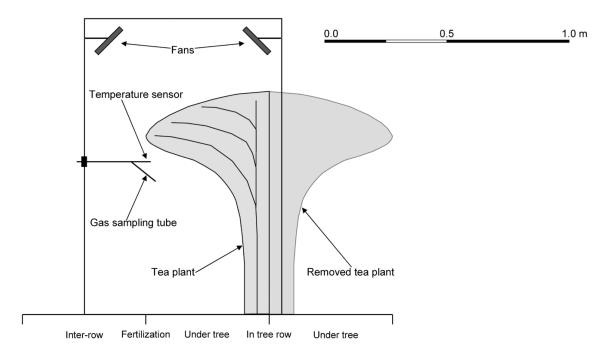


Figure 1. Diagrams of the vertical transect of a tea tree row and the static closed-chamber used for measuring  $N_2O$  emissions from the tea field

Table 1. Descriptive statistics of the  $N_2O$  fluxes ( $N_2O$ , mg  $N/m^2/day$ ), soil water content (SWC, cm $^3/cm^3$ ), soil ammonium-nitrogen content ( $NH_4^+-N$ , mg N/kg soil) and soil nitrate-nitrogen content ( $NO_3^--N$ , mg N/kg soil)

Treatment	Variable	Mean	Minimum	Maximum	CV (%)	Data distribution	
Non-fertilised (CK)	N <sub>2</sub> O	2.10	0.11	8.03	78.6	normal	
	SWC	0.255	0.176	0.317	10.7	normal	
	NH <sub>4</sub> +-N	13.8	0.2	39.2	72.1	normal	
	$NO_3^-$ -N	7.15	1.94	19.7	55.7	normal	
Conventional (CON)	N <sub>2</sub> O	5.03	0.14	18.82	88.4	log-normal	
	SWC	0.283	0.228	0.411	11.5	normal	
	NH <sub>4</sub> +-N	46.2	0.6	227.2	117.1	normal	
	$NO_3^-$ -N	18.9	5.5	42.2	47.8	normal	
Rice straw mulching (SM)	N <sub>2</sub> O	5.04	-1.56	25.80	110.9	log-normal	
	SWC	0.294	0.206	0.392	11.6	normal	
	NH <sub>4</sub> +-N	45.2	2.5	162.4	96.8	normal	
	$NO_3^-$ -N	17.5	5.4	35.7	41.8	normal	

**Soil sampling and analyses**. The soil samples at the depth of 0-20 cm were collected from a 2-3 m diameter circle around chambers during the flux measurements. The soil bulk density was also determined. Each fresh soil sample was homogenised manually and divided into two portions. One portion was used for determining soil volumetric water content (SWC), while the other was analysed for soil ammonium-N (NH $_4^+$ -N) and nitrate-N (NO $_3^-$ -N) using an automated flow injection analyzer (Fiastar 5000, FOSS, Hillerød, Denmark). All of the results of the soil analyses are presented on the oven-dried soil basis ( $105^{\circ}$ C, 24 h). Since July 1, 2010, the SWC was measured by an automatic soil moisture monitoring system (ECH $_2$ O, Decagon, WA, USA).

**Statistical analysis.** The statistical analyses were performed using the R software (http://www.r-project.org). All of the variables were tested for normality and homogeneity of variance. The

Pearson's correlation analysis was carried out to evaluate the relationships between the daily  $N_2O$  fluxes and environmental factors. Differences between management treatments were evaluated by performing t-tests and considered statistically significant at P < 0.05.

## **RESULTS**

**Soil volumetric water content**. For all the three treatments, the SWC data of the 0–20 cm topsoil were normally distributed (Table 1). As shown in Figure 2, during the one-year observation period, CK had a relatively lower SWCs, including lower mean, minimum and maximum values, than those of other fertilised treatments. SM had lower maximum and minimum values than CON, but a higher mean value (0.294 cm³/cm³) than CON (0.283 cm³/cm³).

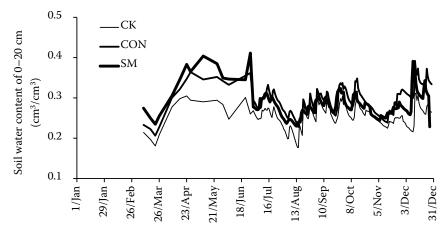


Figure 2. Dynamics of soil water contents  $(cm^3/cm^3)$  of the 0–20 cm topsoil for the three treatments of non-fertilised (CK), conventional (CON) and rice straw mulching (SM) in the tea field

There is a slight discrepancy between the SWCs for CON and SM from March to June, when the SWCs were measured manually, indicating a higher SWC in CON than in SM.

**Soil mineral nitrogen content.** The application of urea on April 15 resulted in a rapid increase on the mineral N content in the CON and SM soils (Figure 3). NH<sub>4</sub><sup>+</sup>-N contents in CON varied greatly, ranging between 0.6 and 227.2 mg N/kg soil, compared with SM (2.5-162.4 mg N/kg soil). As a result of applying the same amount of N fertilisers in CON and SM, two treatments had almost identical NH<sub>4</sub><sup>+</sup>-N mean values (46.2 and 45.2 mg N/kg soil, respectively), which were much higher than that for CK (13.9 mg N/kg soil). The NO<sub>3</sub>-N contents in the three treatments were significantly lower than the NH<sub>4</sub><sup>+</sup>-N contents (P < 0.001) and followed the same pattern as the  $NH_4^+$ -N contents. The soil  $NH_4^+$ -N and  $NO_3^-$ -N contents in CK remained nearly constant during the whole observation year, except for a small increase in  $NO_3^-$ -N during September.

**Nitrous oxide emissions**. The seasonal variations of the  $N_2O$  fluxes are presented in Figure 4. Low fluxes corresponded well to cold seasons, and high fluxes corresponded to rainy seasons. The

N<sub>2</sub>O fluxes in CK ranged from 0.11 to 8.03 mg N/m<sup>2</sup>/day, with an average daily flux of 2.10 mg N/m<sup>2</sup>/day. CON and SM had almost the same average daily fluxes (5.03 and 5.04 mg N/m<sup>2</sup>/day, respectively), but very different ranges of fluxes. The N fertiliser application on April 15 resulted in high N<sub>2</sub>O emissions in CON and SM: the first large emission peak (18.82  $\pm$  4.15 and 25.80  $\pm$  0.28 mg N/m<sup>2</sup>/day, respectively) occurred on May 30 and a second peak (10.85  $\pm$  1.61 and 18.17  $\pm$  0.31 mg N/m<sup>2</sup>/day, respectively) was presented on June 27. The N<sub>2</sub>O fluxes in CK displayed similar N<sub>2</sub>O emission dynamics to those in CON and SM, but they had much lower values during the same period. Using Eq. 1, the annual N2O emissions for the CK, CON and SM treatments were estimated as 7.1, 17.2 and 16.7 kg N/ha/year, respectively. In general, the N<sub>2</sub>O emissions from CK were significantly lower than those from CON and SM (P < 0.01), whereas there was no significant difference between the fluxes from CON and SM (P = 0.82).

Correlation of  $N_2O$  fluxes with environmental factors. The  $N_2O$  fluxes had stronger significant correlations with the daily maximum and minimum air temperatures in SM than in CON (Table 2). The cumulative precipitation of the previous five days

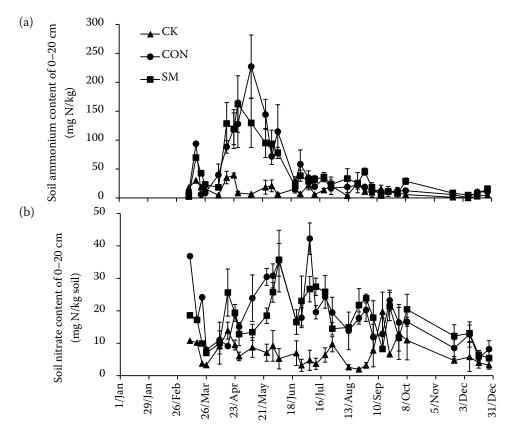


Figure 3. Dynamics of (a) soil ammonium-nitrogen contents (mg N/kg soil) and (b) soil nitrate-nitrogen contents (mg N/kg soil) of the 0-20 cm topsoil for the non-fertilised (CK), conventional (CON) and rice straw mulching (SM) treatments in the tea field. Bars represent standard error

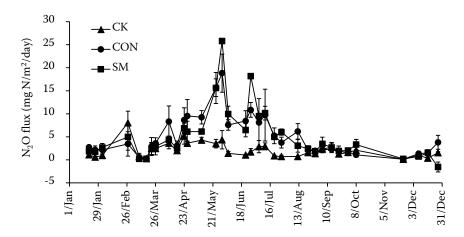


Figure 4. The  $N_2O$  fluxes (mg N/m²/day) from the non-fertilised (CK), conventional (CON) and rice straw mulching (SM) treatments in the tea field. Bars represent standard error

was significantly correlated with the N<sub>2</sub>O fluxes from CON, but not those from SM and CK (r=0.04-0.27, P>0.12), as shown in Table 2. In all three treatments, the N<sub>2</sub>O fluxes were significantly correlated with the SWCs (r=0.45-0.66, P<0.05). Both the soil NH<sub>4</sub><sup>+</sup>-N and NO<sub>3</sub><sup>-</sup>-N contents were significantly correlated with the N<sub>2</sub>O fluxes from CON and SM. However, the NH<sub>4</sub><sup>+</sup>-N contents had a stronger correlation with the N<sub>2</sub>O fluxes than that with the NO<sub>3</sub><sup>-</sup>-N contents in CON (r=0.55, P<0.01 versus r=0.40, P<0.05), while the NO<sub>3</sub><sup>-</sup>-N contents had a stronger correlation with the N<sub>2</sub>O fluxes than that with the NH<sub>4</sub><sup>+</sup>-N contents in SM (r=0.49, P<0.01 versus r=0.39, P<0.05).

## DISCUSSION

N<sub>2</sub>O emissions from fields under different management practices. The N<sub>2</sub>O emissions from soils are mainly attributed to two soil microbial processes, nitrification and denitrification, the most important pathways of N cycling in agroecosystems (Granli and Bøckman 1994). Because  $\mathrm{NH_4^+}\text{-N}$  and  $\mathrm{NO_3^-}\text{-N}$  are the direct substrates for these processes, the more N fertiliser that is applied, the more N<sub>2</sub>O that is emitted from agricultural soils. Our tea field results demonstrate that N fertilisation significantly enhanced  $N_2O$ emissions. As a result of the application of urea, the CON and SM fields emitted more N<sub>2</sub>O than the CK field, and presented several peaks in N<sub>2</sub>O emissions. The high N<sub>2</sub>O fluxes after fertilisation in CON and SM lasted for three months, from mid-April to the end of July, much longer than the effect observed in other field studies (e.g., Liu et al. 2011). The annual N<sub>2</sub>O emissions from CK, CON and SM ranged from 7.1 to 17.2 kg N/ha/year, which are not as high as those from a Japanese tea field (24.3  $\pm$  16.3 kg N/ha for 209 days) reported by Akiyama et al. (2006); however, these tea field  $\rm N_2O$  emissions were markedly higher than those from other land-use types, such as grasslands, forests and paddy fields (Corre et al. 1999). By considering the CK  $\rm N_2O$  emissions as the background emissions and also taking into account the extra 50 kg N/ha/year input from the rice straw mulching, the  $\rm N_2O$  emission factors for the CON and SM treatments were estimated as 2.23% and 1.91% of the total N applied, respectively, for the tea field soils, about twice the IPCC recommended default value (1% of the total N fertiliser input).

Effects of environmental factors on  $N_2O$  emissions. Temperature plays an important role in regulat-

Table 2. Pearson's correlation coefficients between the  $N_2O$  fluxes of the three treatments and environmental factors, including daily maximum ( $T_{max}$ , °C) and minimum ( $T_{min}$ , °C) air temperatures, cumulative precipitation of the previous five days (Rain5, mm), soil water content (SWC, cm³/cm³), soil ammonium-nitrogen content (NH $_4^+$ -N, mg N/kg soil) and nitrate-nitrogen content (NO $_3^-$ -N, mg N/kg soil)

Treatment	$T_{max}$	$T_{\min}$	Rain5	SWC	NH <sub>4</sub> +-N	NO <sub>3</sub> -N
Non-fertilised (CK)	0.04	0.09	0.05	0.45*	0.26	0.20
Conventional (CON)	0.35*	0.34*	0.36*	0.66***	0.55**	0.40*
Rice straw mulching (SM)	0.43*	0.41*	0.27	0.45*	0.39*	0.49**

 $<sup>^*</sup>P < 0.05; ^{**}P < 0.01; ^{***}P < 0.001$ 

ing soil  $\rm N_2O$  emissions. The results of the Pearson's correlation analysis demonstrate that the daily maximum and minimum air temperatures had a positive correlation with the  $\rm N_2O$  fluxes from CON and SM.

Numerous studies reported that soil moisture, N fertilisation and the SOC content are the major driving forces behind  $N_2O$  emissions from agricultural soils (Bouwman 1996). In the CON and SM treatments of this study, the  $N_2O$  emissions were significantly correlated with the SWCs. Keller and Reiner (1994) found the same phenomenon in an old growth forest with relatively high soil nitrate content. However, when there was no N fertiliser applied to CK, only the SWCs were significantly correlated with the  $N_2O$  fluxes (r=0.45, P<0.05).

Hayatsu and Kosuge (1993) reported that acidtolerant or acidophilic autotrophic nitrification can occur in acid tea soils in Japan and that the rate of nitrification increased with an increase in NH<sub>4</sub><sup>+</sup>-N contents from 50 to 300 mg N/kg soil. Our data showed that both the  $NH_4^+$ -N and  $NO_3^-$ -N contents were significantly correlated with the  $\mathrm{N}_2\mathrm{O}$  fluxes in the fertilised tea soils (Table 2). The  $NH_4^+$ -N content exhibited a slightly stronger association with the N<sub>2</sub>O fluxes than NO<sub>3</sub>-N in CON did, while the inverse was observed in SM. This discrepancy implies that nitrification might play a more important role than denitrification in contributing to the N<sub>2</sub>O emissions from CON, and the opposite for SM. Compared with CON, a large amount of fresh organic carbon was applied in the form of rice straw in SM. This rice straw application in SM can change the topsoil micro-environment in tea plantation by decreasing the topsoil temperature, preventing the evaporation and consequently increasing the SWC, and increasing the soil microbial biomass, resulting in more favourable conditions for denitrification. Rice straw mulching may have also improved the N fertiliser use efficiency of tea plantation, similar to the findings of Patra et al. (1993), and thus there were lower annual  $N_2O$  emissions observed from SM than from CON.

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